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Concept Paper

Technologies for decentralized fluoride removal: Testing metallic iron based filters

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Abstract: Since the realization in the 1930s that elevated fluoride concentration in drinking water can have detrimental effects on human health, new methods have been progressively developed in order to reduce fluoride to acceptable levels. In the developing world the necessity for filtration media that is both low cost and sourced from locally available materials has resulted in the widespread use of bone char. Since the early 1990s metallic iron (Fe⁰) has received widespread use as both an adsorbent and a reducing agent for the removal of a wide range of aqueous contaminant species from water. The ion-

selectivity of Fe^0 is dictated by the positively charged surface of iron (hydr)oxides at circumneutral pH. This suggests that Fe^0 could potentially be applied as suitable filter media for the negatively charged fluoride ion. This communication seeks to demonstrate from a theoretical basis and using empirical data from the literature the suitability of Fe^0 filters for fluoride removal. The work concludes that Fe^0 -bearing materials, such as steel wool, hold good promise as a low-cost, readily available and highly effective decentralized fluoride treatment materials.

Keywords: Fluoride removal; Frugal innovation; Packed-bed filters; Steel wool; Water treatment; Zero-valent iron.

1. Introduction

The realization that elevated fluoride levels in drinking water can cause tooth enamel degradation (via dental fluorosis) dates back to 1931 [1]. This discovery subsequently obliged two US cities (Bauxite/Arkansas and Oakley/Idaho) to discard abundant water supplies in favor of those containing lower fluoride concentrations [2]. However, for some localities it became apparent that it is extremely technically difficult to find alternative water sources which contain low fluorine levels [2-4]. To date, countries including China, Estonia, Ethiopia, Ghana, India, Kenya, [Rwanda](#), South Africa, Tanzania and Thailand lack regions which have fluoride below acceptable concentrations for drinking [5-13]. Accordingly, removing excess fluoride from drinking water is a scientific challenge with worldwide significance [14-17].

Boruff [18] was among the first researchers who developed methodology for the removal of excess fluoride from drinking water [2-4,19]. Amongst other materials aluminum oxides and hydroxides received widespread attention [13]. As an example, activated alumina (AA) is synthesized by calcination ($\sim 500^\circ\text{C}$) of aluminum ore (bauxite) to render it porous and highly adsorptive. The fundamental interfacial properties of AA are the protonation of surface hydroxyl groups resulting in the development of surface charge [13,20]. At pH values < 6.0 the surface of AA is positively charged and the adsorption capacity for fluoride is high. In the neutral pH range the affinity of AA surface for fluoride is much lower. This means that, as far as safe drinking water provision is concerned, AA is used below its capacity. Attempts to optimize the AA efficiency for safe drinking water provision include the development of oxides mixtures (Fe_2O_3 , MgO , MnO_2 , TiO_2) [21,22]. These operations appear to have been successful, but the mechanisms are still unknown [13].

In his pioneering work, Boruff [18] tested aluminum sulfate, sodium aluminate, zeolite, activated alumina and bauxite as Al-based materials. He also investigated the suitability of common adsorbents including silica gel, sodium silicate and ferric salts. Despite 80 years intensive research since then, investigations into natural adsorbing materials that can be directly used without modifications (no chemical, physical or thermal treatment) and that are readily available and affordable in the developing world is lacking [14,17,23-25]. In the absence of such natural materials, an alternative is to test and use any manufactured material that is abundantly available and potentially affordable. Because

manufactured materials of concern were not designed for water treatment, pre-treatment (e.g. degreasing) is typically required.

This work aims to establish from a theoretic basis and from using empirical data published in the literature the potential use of Fe^0 for the removal of fluoride (F^-) from drinking water. Fe^0 filters were introduced around 25 years ago for groundwater remediation [26–37] and have received interest for both wastewater treatment [38] and safe drinking water provision [39,40]. However, Fe^0 has received very little attention as filter media for the fluoride removal. In fact, only two peer-reviewed articles on fluoride removal by Fe^0 could be found in the literature: Jahin [41] used nanoscale Fe^0 in batch systems to remove excessive fluoride from water and Jeong et al. [42] treated industrial wastewater in Fe^0 /sand columns operating with 600 V ([external power supply](#)).

Herein an overview of contemporary technologies for fluoride removal from waters is first given before presenting Fe^0 as a new potential technology for decentralized defluoridation in the developing world.

2. Existing technology for F^- removal and their application in the developing world

2.1. Water quality monitoring

The suitability of a water source for any use (including drinking) should be accessed by laboratory analysis in order to determine its biological, chemical and physical composition. Accordingly, the first step to enable (universal) access to drinking water is to equip (and accredit) analytical laboratories worldwide and make them accessible. Based on analytical results there are three options: (i) do nothing because the water source is good for human consumption, (ii) mix contaminated water with clean water (dilution) in order to meet the guideline requirements, and (iii) treat the contaminated water.

In the developed world, district and/or province councils permanently monitor the biological and chemical properties of drinking water and make the results available to citizens. Citizens who are dependent on private water supply have the possibility to make water analysis in accredited laboratories. Moreover, private suppliers have the obligation to monitor the quality of supplied water [43]. In the developing world, analytical laboratories are scarce and where available, water analysis is not affordable for the majority of the population.

This presentation underlines this discrepancy. In the era of modern analytical instrumentation, this discrepancy is not acceptable. Accordingly, a mandatory pre-requisite to universal access to safe drinking water is to equip water laboratories in the developing world [40].

2.2. Water treatment technologies for fluoride (F^-) removal

Conventional F^- treatment methods, which are typically applied in centralized treatment plants, include: reverse osmosis, electrodialysis, coagulation-flocculation-filtration and adsorption with activated alumina [9,13,25]. In contrast the most popular alternative (often decentralized) F^- treatment methods are: (i) adsorption with local materials (clays and soils, brick chips), (ii) adsorption with activated alumina and bone char, and (iii) the Nalgonda technique [5,8,24,25,44,45]. Water filtration on biomaterials and bone char will be presented in the next section as common absorbents in the developing world (Tab. 1).

Table 1. Availability of common adsorbents for aqueous fluoride removal.

Material	Comments	Availability
Metal oxides and hydroxides	Exhibit relatively low efficiency without physico-chemical modification	low
Biosorbents	Exhibit relatively low efficiency without physico-chemical modification	very high
Geomaterials	Exhibit relatively low efficiency without physico-chemical modification	high
Carbonaceous Materials	Typically require expensive physico-chemical activation	low
Bone char	Produced by carbonizing animal bones	high
Industrial by-Products	Typically require expensive physico-chemical activation	low
Metallic iron (Fe^0)	Under natural conditions a ubiquitous (hydr)oxide layer is also present	high

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2.3. Biomaterial based filters for F^- removal

There is a growing interest in the application of plant biomass for aqueous F^- removal [46-51]. The main advantage of plant biomass is that it is readily available and can operate at circumneutral pH. The F^- removal mechanism by this class of material is not well established. It is currently understood that the driving force for F^- removal relies in its affinity to several species present in the biomass structure (e.g. Ca^{2+} , Mg^{2+} , OH^- and $-\text{NH}_2$ groups) [47,48,52,53]. Several plant materials such as bark from *Moringa olifera* and *Embllica officinalis*, the roots from *Vetiveria zizanoides* and the leaves from *Cyanodon tactylon* sisal have been determined as effective defluoridating agents ([51] and refs. cited therein). Literature reports that tea leaves could quantitatively remove fluoride from a aqueous 20 mg/L solution [54], Tamarind (*Tamarindus indica*) fruit cover powder treated with HCl, could remove up to 57.1 % of fluoride from a natural water containing 3.5 mg/L F^- at the pH of 7.6 [46]; Comparative study using a synthetic solution containing fluoride at 2.0 mg/L, showed that, biomass from vetiver (*Vetiveria zizanoides*) roots, tamarind seed (*Tamarindus indica*), clove (*Eugenia carryophyllata*), neem (*Azardirachta indica*), acacia (*Acacia catechu* willd), nutmeg (*Myristica fragrans*) and coffee husk (*Coffea arabica*), could remove up to 80 %, 75 %, 70 %, 52 %, 47 %, 45 % and 38 % of fluoride ion respectively, at neutral pH [48]. A more recent comparative study reported F^- removal efficiencies of maize leaf, goose grass, banana false stem, untreated sisal fibre and Aloe vera leaf and sisal pith biomass to be 4.1, 4.6, 7.1, 26.6, 29.4 and 47.3 % respectively.

2.4. Bone char based filters for F^- removal

Bone char (BC) is generally considered as the most efficient and affordable material for F^- removal in decentralized safe drinking water production at present [25,55-58]. The four main reasons for this are: (i) BC is relatively inexpensive, (ii) BC is made from locally abundantly available spent bone, (iii) BC

can be manufactured in sufficient amounts and quality for local needs and (iv) BC can be recycled via a simple thermal treatment process at 200 °C.

BC predominantly consists of hydroxyapatite $[\text{Ca}_{10}(\text{PO}_4)_6(\text{OH})_2]$ [2,58-60]. BC therefore retains fluoride by chemisorption and ion-exchange between dissolved F^- ions and PO_4^{3-} . Electrostatic attraction of dissolved F^- is also known to occur in parallel with F^- adsorbed via physisorption. The dominating mechanism of fluoride removal has therefore not been fully elucidated [58,61,62], however, it is certain that the kinetics of hydroxyapatite dissolution limits the extent of ion-exchange.

BC has been used as a versatile adsorbent for a wide variety of other pollutants, including arsenate, dyes and heavy metals [58]. Therefore, BC has the potential to be used universally for decentralized safe drinking water provision in chemical-free systems. However, some attempts to use locally produced BC for aqueous F^- removal have not been successful [8]. Two major reasons for this are: (i) a societal aversion exists for the handling of fresh bone; and (ii) BC filtered water produces an unpleasant taste [11,58]. Accordingly, alternatives to BC are still sought. Fe^0 materials in packed-bed are a potential candidate.

3. Fe^0 filters

Fe^0 -based filters have been demonstrated as able to remove/inactivate different types of biological and chemical pollutants from water [63-71]. The performance of individual filters depends on the Fe^0 type (intrinsic reactivity), Fe^0 particle size (mm, μm or nm scale), filter design (e.g. depth of the reactive layer, dimensions of the filter), type and concentration of complimentary material (e.g. Fe^0 /sand ratio), water composition (e.g. pH and presences of competing species) and operating conditions (e.g. water flow, temperature). Thus, there is diversity among operational factors that can impact the efficiency of a Fe^0 filter. This high degree of diversity for significant factors suggests that only a well-designed systematic approach could identify conditions for optimal operation [40,72-76].

3.1. Historical background

During the past two decades, great efforts have been made to develop Fe^0 filters for use in decentralized water supply systems in the developing countries, including Argentina, Bangladesh, Cambodia, [Cameroon](#), Chile, India, Nepal and Vietnam [39,40,63,65,71,77-85]. Several Fe^0 filters exist and were also developed/optimized by scientists from many institutions worldwide, including the Massachusetts Institute of Technology [77,78], the George Mason University [80] or the Swiss Federal Institute of Aquatic Science and Technology [85,86].

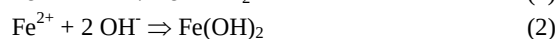
Reviewing the history of Fe^0 filter development, Noubactep et al. [75] demonstrated that with exception of the SONO filters [39,80], the volumetric expansive nature of iron corrosion [87,88] was not properly considered. Incorporating the volumetric expansion of Fe^0 can be achieved via three ways: (i) decreasing the proportion of Fe^0 in the filter [76,89], (ii) using complimentary porous materials [39,80] and/or (iii) limiting the dissolved O_2 concentration in the contaminated water [40,89,90]. Another limitation of the early Fe^0 filters was that they were designed for Fe^0 to serve either (i) as electron donor for contaminant reduction [91], (ii) as iron oxide generator for species

removal by adsorption and co-precipitation [92] or (iii) as iron oxide generator to sustain sand filtration depending on the species to be removed [93-100].

3.2. The state-of-the art

A central rationale for the applicability of Fe^0 materials for water treatment is due to their ability to produce solid iron (hydr)oxides corrosion products which are highly effective aqueous contaminant scavenging agents [101-103] for negatively charged substances [45,95,104,105], such as F^- [40].

Another important feature of the $\text{Fe}^0/\text{H}_2\text{O}$ system is that both Fe^0 oxidative dissolution (Eq. 1) and the precipitation of Fe oxide-hydroxides (Eq. 2) occur in the presence of contaminants. This makes interactions between dissolved species, Fe^0 , Fe oxide-hydroxides and intermediate species very complex [106,107]. Intermediate species include Fe^{II} and H/H_2 [108] which are reducing agents, with increased reducing power when already adsorbed onto (nascent) oxides [109,110]. The complexity of processes in $\text{Fe}^0/\text{H}_2\text{O}$ systems has led to the misconception, that Fe^0 is a lone reducing agent [26,29,33,36,111,112]. This may explain why Fe^0 has not been widely tested for F^- removal, i.e. the standard electrode potentials for fluoride ($E^0 = 2.87 \text{ V}$ - Eq. 3) and Fe^0 ($E^0 = 0.44 \text{ V}$ - Eq. 4) indicate that Fe^0 in theory is unable to reduce F^- (i.e. F_2 would oxidize Fe^0).



There are two other important features of Fe^0 corrosion pathways that are essential for the process of contaminant removal: (i) the volume of each oxide (V_{oxide}) is larger than the volume of the parent metallic iron (V_{iron}) [87] and (ii) during the precipitation of oxides, contaminants can become enmeshed (i.e. physically sequestered) [106,107]. The volumetric expansive nature of iron corrosion ($V_{\text{oxide}} > V_{\text{iron}}$) implies that irrespective from any adsorptive affinity between oxides and pollutants, decontamination by size exclusion is improved with the filter service life. On the other hand, 'pure physical sequestration' can also occur independent of the Fe (hydr)oxide adsorptive affinity [113]. Therefore, $\text{Fe}^0/\text{H}_2\text{O}$ systems are generally considered to be more suitable for negatively charged species but are often effective for the removal of other species as well [104,105,114].

Aqueous contaminants are typically removed in Fe^0 filters by adsorption, co-precipitation and size-exclusion, making Fe^0 filters a reactive filtration as opposed to adsorptive filtration in iron-oxide-coated sand filters. Here, iron oxides are progressively generated in the system [101,115-118], making the efficiency of a Fe^0 filter highly dependent on its physical design (e.g. Fe^0 reactivity, filter bed length, flow velocity). As a sinuous and tortuous material with therefore significant intrinsic porosity (available voids which can be used to accommodate the expansive Fe^0 corrosion products) steel wool (SW) can be considered a suitable candidate for use as filtrate material [72-74,89,199,120].

3.3. Designing Fe^0 filters for fluoride removal

3.3.1. Physico-chemical processes in Fe^0 filters

Several processes take place during Fe^0 filter operation: (i) in-situ generation of Fe corrosion products, (ii) adsorption of dissolved species onto strained solids, and all available surfaces, including Fe^0 and its (hydr)oxides, and (iii) biologically induced processes such as biofouling [35,65,68,93,94,108]. Depending on the Fe^0 type (intrinsic reactivity), the size and depth of the filter, the water flow velocity and the chemical and physical properties of the ‘raw water’, well-designed Fe^0 filters can free water from all possible pollutants, namely: (i) chemical contaminants (organic and inorganic, including tastes and odours), (ii) biological contaminants (pathogens, including bacteria and viruses) and (iii) physical contaminants including suspended solids [40,75,81,82,121].

3.3.2. Biological mediated processes in Fe^0 filters

A Fe^0 filter can be adapted to become a “self-sustaining” filter, such as a slow sand filter (SSF). For over 200 years, SSFs have been used to control microbiological water contamination ([122] and refs. cited therein). There is currently a renewed interest in SSF application as systems operating without chemicals or electricity. The processes involved within biofilms of SSF are still not fully understood. Relevant processes include adsorption, bio-oxidation, predation and scavenging. These processes have been hypothesised but have never been comprehensively verified [122,123]. Such a gap in knowledge pertaining to the optimal efficiency of SSFs hampers advances in their design.

Because Fe^0 filters operate under similar conditions like SSFs, microbiological communities will develop within them and potentially impact the process of iron corrosion as well as the transformation of corrosion products. As an example, crystalline Fe^{III} oxides can be microbiologically reduced to Fe^{II} oxides via colloidal hydroxides [124,125]. Thus beside abiotic processes, biotic processes also contribute to decontamination in Fe^0 filters and should also be considered.

3.3.3. Optimising the efficiency Fe^0 filters

The efficiency of Fe^0 filters can be considerably increased when suspended solids are removed in pre-filters (e.g. filtration on coarse and fine sand). This operation increases the service life of Fe^0 filters. The service life of Fe^0 filters is also considerably lengthened if the formation of more voluminous corrosion products is avoided by working under anoxic conditions [76,88,89,126]. Low oxygen levels are known to be achieved in biofilms of SSFs. In fact, dissolved oxygen is used for respiration and hence the development of an aerobic microbial community building the biofilm [122]. In other words, an ideal Fe^0 filter utilises both the O_2 removing capacity of a SSF and the well-documented aqueous contaminant scavenging ability the $\text{Fe}^0/\text{H}_2\text{O}$ system [76,121,126]. A suitable design may consider the periodic replacement of the SSF units (e.g. after every 6 months) in order to utilise the full contaminant scavenging ability of the Fe^0 (Fig. 1).

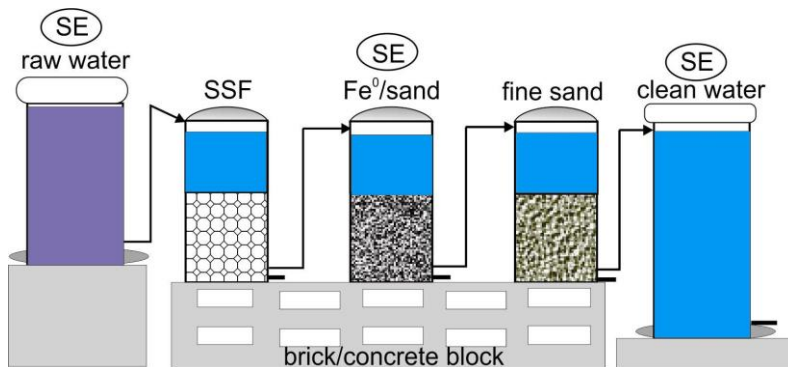


Figure 1: Side view of a sequential Fe^0 -sand water filter. To optimize the efficiency of the filtration unit, the effluent should be clarified and deaerated before being introduced, using for example, a slow sand filter (SSF). Water from the Fe^0 filter subsequently flows through a sand layer. “SE” specifies the positions where solar energy is expected to optimize operation and monitoring of Fe^0 filter. SE may have four key functions: (i) pumping the ‘raw water’ to a storage tank, (ii) regulating the flow regime out of the storage tank, (iii) controlling the temperature of the Fe^0 -based units, and (iv) pasteurizing the stagnant treated water. Figure is reproduced from Salama et al. [127].

In water treatment scenarios where SSFs are too slow to produce low O_2 concentrations (typically < 2.0 mg/L), other O_2 scavengers can be used, including Fe^0 itself [29,89,90]. When Fe^0 is used as a dissolved O_2 scavenger, the user should remember that any preferential flow created in the sacrificial Fe^0 unit will likely impact the efficiency of the treatment unit(s). In this case equalizing units (gravel or anthracite columns) should be used [128,129]. Ideally, materials contained in the equalizing units should have good adsorptive affinity for positively charged species which therefore have less affinity adsorb to iron oxides [104,105]. For the present study, however, equalization is the major goal as fluoride ions are negatively charged and thus possess good adsorptive affinity for the surface of Fe oxide-hydroxides which cover the Fe^0 surface.

3.3.4. Fe^0 filters for fluoride removal

Section 3.3.1 discussed the parameters that determine the efficiency of Fe^0 filters. From the material perspective, three relevant parameters are named: Fe^0 type, Fe^0 mass and Fe^0 particle size. The depth of the reactive zone (e.g. Fe^0 /sand mixture - Fig. 2) in the filter is a further important parameter. The water flow velocity is a parameter that can easily be fixed, for example using the same as employed for the SSF.

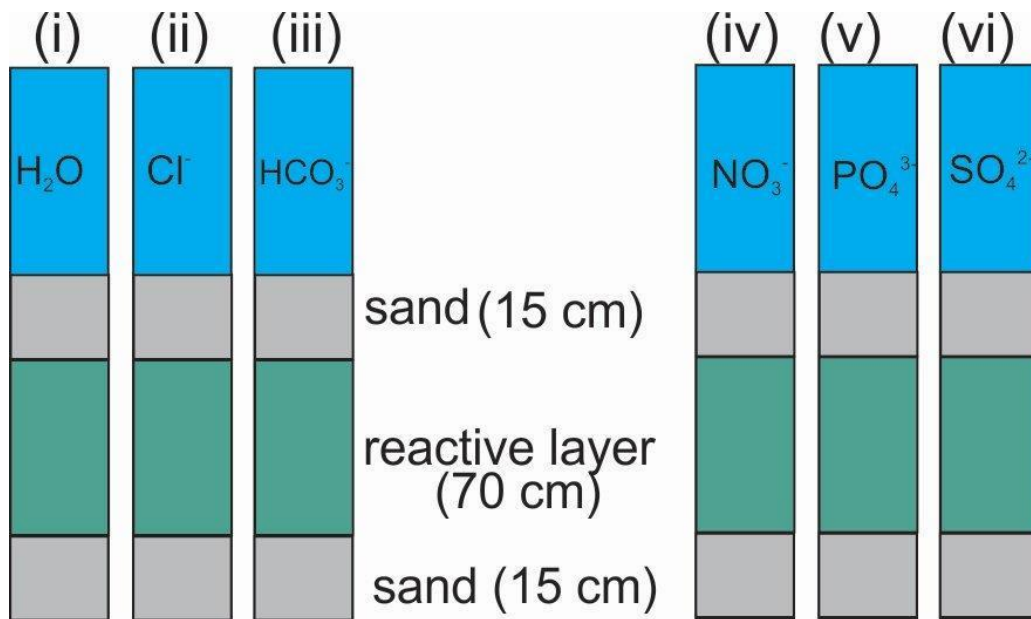


Figure 2: Filtration unit modules used to investigate of the impact of different anions on the performance of a Fe^0 packed bed for F^- removal. Layered arrangement of iron filings and sand used for the experiments is also shown, however, by modifying this design other parameters (e.g. Fe^0 type, sand grading/sorting) can also be investigated.

For aqueous F^- removal in $\text{Fe}^0/\text{H}_2\text{O}$ systems, the pH value is not typically very important as only natural waters with $\text{pH} > 4.5$ are concerned [130]. In this pH range, the solubility of iron is minimal, fluoride is negatively charged and the surface of Fe oxide-hydroxide positively charged. The concentration of dissolved anionic species (e.g. Cl^- , HCO_3^- , NO_3^- , PO_4^{3-} , SO_4^{2-}), added for example as sodium salts, are often a more important factor than simply the pH value (Fig. 3), however, the speciation of these ions (namely HCO_3^- , PO_4^{3-} and SO_4^{2-}) is pH dependent. However, as F^- removal is the main subject, the pH-dependant speciation of co-existing ions is of secondary importance (for a conceptual investigation).

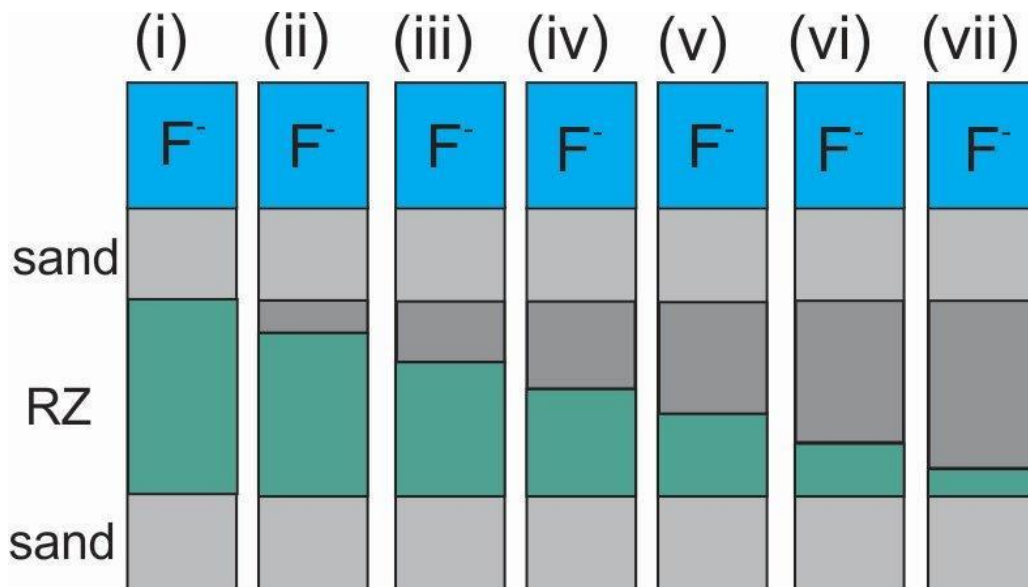


Figure 3: Filtration unit modules used for the investigation of the effects of the thickness of the reactive zone (RZ) on the performance of a Fe^0 packed bed for F^- removal. Layered arrangement of iron filings and sand used for the experiments is roughly shown. Sand in RZ and is not necessarily the same at the inlet/outlet.

In order to investigate the efficacy of a water filtration technology a systematic approach much be adopted which comprises fixing certain parameters and testing all other in a pre-defined sequence. For example, filter container (10 cm inner diameter; 100 cm height), depth of the reactive zone ($H_{rz} = 70$ cm). The other 30 cm could be made up of two times: coarse gravel: 5 cm, river sand: 5 cm and fine gravel (drainage layer) (Fig. 2 and Fig. 3): 5 cm (one at the inlet and one at the outlet). Ideally, twenty five percent of the reactive zone (70 cm) -volumetric ratio- should be made up of Fe^0 . Accordingly, comparatively testing the efficiency of three Fe^0 materials for F^- removal should consist in performing parallel experiments with three systems (ideally each in triplicates) differing solely by the Fe^0 type in the reactive zone (modified Fig. 2; each column contains another Fe^0 material and the influent is the same). Ideally the breakthrough of F^- and Fe are monitored as function of time, as well as the pH value and any changes in the hydraulic conductivity [40,114]. It is understood that the dissolved O_2 level and common water quality parameters will be routinely monitored and documented. Relevant parameters include electric conductivity, temperature, Ca^{2+} , Na^+ , K^+ , Mg^{2+} , Mn^{2+} , Cl^- , HCO_3^- , NO_3^- , PO_4^{3-} , SO_4^{2-} . Based on the results of Fe^0 efficiency characterization, a material with acceptable corrosion kinetics could be selected to test the impact of all other operational parameters. For example, testing the impact of the five named anions (Cl^- , HCO_3^- , NO_3^- , PO_4^{3-} , SO_4^{2-}) should consist of testing six systems in parallel, the sixth system being the fed by the operational solvent (e.g. tap water) (Fig. 2). Ideally, relevant anion concentrations for specific sites should be tested. Alternatively, experiments with equivalent amounts of the anions uncovering the range of natural concentrations could be performed.

Such principle experiments are necessary to gain a general overview of the impact of all relevant influencing parameters and thus, ease site-specific fine-tuning experiments.

The last type of experiments will consist in determining the depth of the reactive zone of filter bed needed for efficient water treatment for a certain period (e.g. 6 months) (Fig. 3). This is to determine the frequency for the replacement of Fe^0 units. This investigation implies the determination of the numbers of beds having a reactive zone of 70 cm that will be needed to efficiently treat water for 6 months.

The next section will discuss methods to test commercially available SW [116–118,131,132] as a suitable fluoride filtration technology. If this operation is successful, such filtration systems will be able to be manufactured in the developing world using commonly available tools [39,77,78,80,84,133,134].

4. Steel wool for universal fluoride removal

4.1. Availability and suitability of steel wool

Metal wools are globally abundant. They are used in sanitation and maintenance for cleaning and polishing many types of surfaces including granite, marble, metal and wood. The three most common types of metal wools are steel (Fig. 4), bronze and copper. SW is often made of various raw materials including low alloyed steel and stainless steel. For the purpose of water treatment a rule of thumb could be: “use only the SW that readily rusts in water” (i.e. no stainless steel wool).

The thickness of SW fibers (μm range) suggests that they can deplete within a year or some two years upon aqueous corrosion [135,136]. This makes SW a practical material to start characterization of key aspects of the long-term Fe^0 efficiency (sustained reactivity), both at lab and field scales. Long-term parallel experiments with SW (e.g. diameters 30 to 300 μm – ref. [135]) and thicker Fe^0 materials (e.g. shopped iron wires of different diameters $> 300 \mu\text{m}$) are recommended in this context.

Whilst SW is an extremely widely used and cheap material some commercial forms of SW contain oil and grease from the manufacturing process that should first be washed (cleaned) before use [69,137,138]. In addition a mild acid treatment may be necessary in order to remove any surface contamination including Fe^{III} bearing (hydr)oxide phases.



Figure 4: Photographs of six steel wool specimens found on the market in Douala/Cameroon (August 2015).

4.2. Current use of steel wool for water treatment

Tseng et al. [137] used SW as generator of iron oxides for the ‘accumulation’ of ^{60}Co (pre-concentration) in environmental monitoring of a nuclear power plant. This demonstrates that this material is a potential suitable candidate to be used for water treatment in Fe^0 filters. Since then, James et al. [101] tested SW as an amendment material for peat and sand beds in order to: (i) increase the performance for phosphate removal and (ii) lengthen the service life of such beds. The evidence that SW amended peat/sand beds are sustainable systems for phosphate removal was subsequently described some 15 years later by Erickson and colleagues [116–118].

With the introduction of the Fe^0 technology, several investigators have tested various SW materials for water treatment (Tab. 2). Applications included in-situ production of (i) H_2 for denitrification (Biswas and Bose [145] – Nr. 11), (ii) Fenton reagents for oxidative degradation (Teixeira et al. [141] – Nr. 1), (iii) iron oxides for As removal (Campos [132] – Nr. 12) and (iv) iron oxides for pathogen removal (Bradley et al. [95] – Nr. 6). However, investigations were performed on a pragmatic basis (Tab. 3 and Tab. 4) and have not considered pioneering works (Nr. 15 and Nr. 16, Tab.2). The used experimental conditions were not properly documented (Tab. 2) and were so diverse (Tab. 3 and Tab. 4) that available results have solely a qualitative value (section 4.3). Additionally, precedent works testing

SW for decontamination in the framework of the ‘Fe⁰ remediation technology’ have not been considered as a rule. Table 5 shows clearly that from the six most recent studies using SW, just one (1) of seven (7) possible previous peer-reviewed articles is referenced in two (2) from six (6) articles. This evidences the fact that SW is just qualitatively used as Fe⁰-based materials as it has been the case in classroom demonstrations of the oxygen content of air (approximately 1/5 of the dry air) for several decades [146–149]. For example, Birk et al. [146] allowed acid-washed steel wool to rust in an air-filled container inverted in water; the decrease of air volume is measured. In this context, the experimental vessels can be heated and all other possible tools can be used to accelerate Fe⁰ oxidation [148]. The intrinsic reactivity of used SW is not addressed; just its capability to reduce atmospheric O₂ is of interest.

Summarized, 30 years ago Tseng et al. [137] endeavored to use SW as source of iron oxides for aqueous contaminant removal. Since 1997, testing/using SW has become a common tool for the removal of individual aqueous species (Tab. 2, 3 and 4). Knowledge on this process is relatively poorly reported in the scientific literature at present, i.e. a systematic approach to investigate the operating mode of SW for water treatment is currently lacking. Considering the shortcoming of previous works, Section 4.4 introduces a concept for testing SW based Fe⁰ filters for fluoride removal at household level. The relevance of data/parameters summarized in Tab. 2 through Tab. 5 is discussed in the next section.

Table 2: Summary of trade names, iron and carbon contents of fourteen peer-reviewed articles using steel wool for contaminant removal. It is seen that only four articles have specified the composition of used steel wools while seven have not specified the trade name of their used specimens. ‘Nr’ is the number referencing individual articles in Tab. 3, 4 and 5.

Nr	Anno	Ref.	Trade name	Element (%)	
				Fe	C
1	2015	[141]	Limpano (Brazil)	99.8	0.1
2	2014	[140]	n.s.	n.s.	n.s.
3	2013	[139]	n.s.	n.s.	n.s.
4	2013	[136]	n.s.	n.s.	n.s.
5	2012	[138]	n.s.	98.6	n.s.
6	2011	[95]	Peerless Metal Powders & Abrasive	n.s.	n.s.
7	2010	[142]	Bombril (Brazil)	n.s.	n.s.
8	2009	[143]	Mapavirulana(Argentina)	98.5	0.1
9	2008	[144]	n.s.	98.5	n.s.
10	2007	[117]	Global Material Technologies (USA)	n.s.	n.s.
11	2005	[145]	Rohit Industries (India)	n.s.	n.s.
12	2002	[132]	n.s.	n.s.	n.s.
13	1998	[131]	Rhodes American Company	n.s.	n.s.
14	1997	[135]	n.s.	n.s.	n.s.
15	1992	[101]	Rhodes American Company	n.s.	n.s.
16	1984	[137]	Rhodes Company (USA)	n.s.	n.s.

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Table 3: Summary of some experimental conditions used for the batch experiments in thirteen of the peer-reviewed articles using steel wool for contaminant removal. The numbers (Nr.) are related to relevant references as specified in Tab. 2. Polychlorinated biphenyls are abbreviated to: “PCBs”.

Nr.	X	[X] (mg/L)	Fe ⁰ (g/L)	pH value	Duration	Volume (mL)	Mixing	Speed (rpm)
1	Phenol	200	1 to 7	5 - 9	300 min	500	stirred	n.s.
3	Cr	100	0.0083	3	2 h	50	n.s.	-
4	As(V)	0.3	n.s.	3 -12	480 min	100	shaken	35-40
5	PCBs	< 12800	500	n.s.	4 h	2	stirred	n.s.
7	Cr(VI)	9.44	0.015	n.s.	3 h	30	n.s.	n.s.
8	As	1	2 to 3	7.1	150 min	500	stirred	300
9	As	1 - 1.3	1.3	< 8	24 h	1000	quiescent	-
10	PO ₄ ³⁻	0.5	n.s.	5.7	24 h	150	shaken	-
11	NO ₃ ⁻	42	2.5 - 25	n.s.	60 d	200	n.s.	n.s.
12	As	0.5 to 10	n.s.	6.1 to 6.8	15 min	50	n.s.	n.s.
13	NO ₃ ⁻	50	40	5 to 11	12 d	250	shaken	100
15	PO ₄ ³⁻	4	7.41	5	60 min	25	agitated	-
16	⁶⁰ Co	200-1000	10	6.84	30 s	20	n.s.	n.s.

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Table 4: Summary of some experimental conditions used for the column experiments in nine of the peer-reviewed articles using steel wool for contaminant removal. The numbers (Nr.) are related to relevant references as specified in Tab. 2. Point forming unit is abbreviated to: “PFU”.

Nr	X	[X]	Fe ⁰	Dimensions (cm)			Porosity	pH	Duration	Flow
		(mg/L)		ID	L	RZ				
2	Cr	< 100	10	3	30	n.s.	n.s.	3–7	120 min	20
4	As (V)	0.3	47	4	0.1	n.s.	n.s.	n.s.	90 h	< 600
5	Virus	10 ¹⁰ PFU	260	2.5	8.9	n.s.	n.s.	6.2	170 d	< 38.4
9	PO ₄ ³⁻	< 0.8	< 5.5	5.08	84	n.s.	n.s.	n.s.	< 2 d	n.s.
10	NO ₃ ⁻	40	< 1.5	4	n.s.	12	50	n.s.	n.s.	n.s.
13	NO ₃ ⁻	50	8	2.5	26.5	n.s.	90	6.7	131 d	3 and 10
14	Cr(VI)	500	1	3.5	20	n.s.	n.s.	n.s.	n.s.	15000
15	PO ₄ ³⁻	50	0.3	n.s.	< 2.9	n.s.	n.s.	n.s.	22 d	4.2
16	⁶⁰ Co	<180	40	5	n.s.	7	n.s.	6.84	< 24 h	72000

Table 5: Extent to which recent articles which have investigated using steel wool for water treatment were considered by academic journal papers since 2010 (Nr. 1 through 6, Tab. 2). It is evident that existing articles are typically independent work as just one ancient article was cited by 2 of the 6 articles since 2010. Additionally none of the pioneering work (Nr. 15 and 16) was considered by all articles of the Fe⁰ remediation technology area.

Nr	Recent articles						
	7	8	9	10	11	12	13
1	no	no	no	no	no	no	no
2	no	no	no	no	no	no	yes
3	no	no	no	no	no	no	yes
4	no	no	no	no	no	no	no
5	no	no	no	no	no	no	no
6	no	no	no	no	no	no	no

4.3. Lessons from the peer-reviewed articles on ‘SW for water treatment’

Tables 2 through 5 include peer-reviewed articles only. Academic thesis [116], technical reports [150,151] and patents [152] do exist as well but are not considered herein because they have also not specified key issues addressed herein. This section will corroborate the argumentation from Section 4.2 by giving a short comment on each table.

Table 2 shows that 6 from 14 articles have not specified the trade name of the materials they used. Given the trade name is not necessarily a must but given the elemental composition of used materials should be mandatory. Only four articles have done this. Additionally, only three articles have specified the specific surface area (not shown) of the material they tested. Accordingly it is fair to state that efforts to define used steel wools were not satisfactorily.

Table 3 summarizes the experimental conditions for the 13 articles testing SW in batch experiments. It is seen that seven different contaminants (As, Co, Cr, NO₃⁻, PCB, phenol and PO₄³⁻) were tested. The initial concentrations varied from 0.3 to 12800 mg/L. The SW mass loading, where specified, varied from 0.01 to 500 g/L. The initial pH value, if specified, varied from 3.0 to 12.0 (for the purpose herein only pH > 5.5 is relevant). The experimental duration varied from 30 s to 60 days. The volume of the reaction vessels varied from 20 to 1000 mL and several mixing speeds and intensities were tested. The suitability of these operational parameters to achieve the given objectives of the individual studies has been discussed already and will not be repeated here [35,153]. It is sufficient here to recognize that results obtained under such diverse experimental conditions are highly qualitative. Additionally batch experiments are of little usefulness when it comes to designing steady-state flow systems.

Table 4 summarizes the experimental conditions for the 9 articles testing SW in column experiments. Six different contaminants (As, Co, Cr, NO₃⁻, PO₄³⁻ and a virus) were tested. The initial concentrations varied from 0.3 to 500 mg/L. The SW mass varied from 1.0 to 260 g/L. The dimensions of the columns and the height of the reactive zone (SW-containing layer) were not always specified. The initial pH value, where specified, varied from 3.0 to 7.0. The experimental duration, where specified, varied from

2 h to 170 days. The flow velocity, where specified, varied from 4 mL/h to 15 L/h (72 L/h is not considered as it was used for pre-concentration). Again the large variability of tested experimental designs degrades available column results to qualitative results, even for columns testing the same contaminant like Nr. 2 and Nr. 14 (testing SW for Cr removal). This situation is not ideal as lab scale columns experiments have the potential to be used as bench scale tests for household filters. For example, testing the approach if Biswas and Bose [145], given the mass of SW in 125 cm³ of sand, one can rapidly select SW-materials for further test and optimize the operational parameters for efficient Fe⁰-SW filters. The question to answer in this primarily research is twofold: (i) which SW samples are suitable for water treatment and (ii) which mass of suitable material per 125 cm³ should be used? Adopting such an approach within and across research groups will certainly accelerate knowledge acquisition for efficient SW-based Fe⁰ filters.

Table 5 compares the extent at which post-2010 peer-reviewed articles have referenced previous articles (1992 < year ≤ 2010) dealing with SW for water treatment. The preponderance of ‘no’ in Tab. 5 (40/42) clearly demonstrates that available works are independent efforts. The scientific community cannot afford this approach anymore. In essence, presenting new results without putting them in an historical context is not acceptable and even not fair for reader wishing at learning from a given research field.

4.4. Steel wool based Fe⁰ filters for water treatment at household level

A suitable example design of a stand-alone SW filtration system is described herein. SW was chosen as a suitable Fe⁰ material because of its wide commercial availability at relatively low cost and its high surface area with little resistance to fluid flow [101,117,141]. Due to the high levels of variability of both steel wool (SW) samples and contaminated water sources, only a systematic sequence of experiments can determine whether a SW specimen is suitable for water defluoridation.

Laboratory based studies should be conducted to investigate the effect of bed design variables on the extent of F⁻ removal, mostly in gravity driven system. Design parameters to be investigated include: (i) SW type, form and quantity, (ii) supporting filter media (e.g. anthracite, pumice, sand) and their particle sizes (e.g. 0.5 mm, 1.0 mm, 2.0 mm), (iii) dimension of the filter (e.g. ID = 10 mm and L = 50 mm), (iv) depth of the reactive zone (e.g. 15 mm, 30 mm or 40 mm) (Fig. 3), and (v) packing configurations of the reactive zone (layered or mixed).

General rules of thumb for filter design can often be used as a guideline. For example: (i) the size of filter media particles significantly impact the hydraulic conductivity (permeability) and thus the overall contaminant removal performance, (ii) filters with smaller particles had better straining efficiency, but subsequently shorter lifespan, (iii) deeper filters have longer lifespan, (iv) two layers of distinct sized media in the filter bed improved performance over the single-layered systems, (v) mixed systems showed improved performance, as compared with single-layered systems and (vi) low water flow velocities are better for filter performance [40,76,154,155]. It is certain that simple modifications of SW based water filters can significantly affect their performance and/or reduce maintenance intervals.

4.5. Proposed design procedure

The idea of using Fe^0 filtration for universal safe drinking water provision [72] has now progressed to the point where point-of-use systems should be piloted [40]. Although no real world prototype has been presented yet, examples given below target at designing typical SW based systems for individual households. This option may be improved in future and tools presented in related articles [40,67,68,75,121] would assist in scaling up for larger population sizes.

4.5.1. Required water production rate

It is assumed that an average household (model household) in a rural settlement in the developing world has ten inhabitants. The World Health Organisation (WHO) recommends 7.5 litres (L) of water per person per day, meaning that the model household needs a minimum of 75 L of water per day. Considering a surplus water use of 25 L per day a suitable filtration system would therefore be required to produce approximately 100 L of water per day. Furthermore the system could be designed such that 50 L of water is filtered in the morning and the remaining 50 L in the evening.

4.5.2. Media and water properties

Locally available anthracite, gravel, pumice or sand should be (crushed and) sieved and the most suitable particle sized (e.g. $0.5 \leq d(\text{mm}) \leq 1.5$) used [154]. Fe^0 material herein is operationally SW that is either used as received or chopped into pieces of suitable sizes. SW should be characterized for its chemical composition, density, chemical reactivity after available tests ([156] and referenced cited therein). The 'raw water' quality has to be analytically determined and after the filter effluent must contain F^- at less than 1.5 mg/L (WHO guideline – [157]).

The major challenge of this research is to experimentally determine a suitable media volume (or mass) required to efficiently treat 100 L water/day at a minimal filtration rate of 5 L/h (50 L in 10 h). The maximal filtration rate should also be specified as well as the suitable bed depth, the mass of SW in a predetermined volume of the reactive zone. Once this goal is achieved the service life of such a filter should be specified and ways and means should be sought to lengthen it to at least 6 months (if applicable). Relevant tools include using several beds in series [40,158].

First exploratory research could start with a reactive zone of 500 mL made up of graded sand (e.g. $0.5 \leq d(\text{mm}) \leq 1.5$) and 10 % (w/w) chopped SW. Depending from the results, the SW proportion can be varied as well as the depth of the reactive zone. It is understood that for each SW sample the experiment has to be started at point zero until some correlations are established between chemical composition, density and efficiency at removing F^- . It is also understood that the reactive zone is confined by two layers of granular material with incidence on the water quality and the water flow velocity. It has already been discussed that the layering of materials at the outlet should be selected to avoid high iron concentration in treated water [40,75].

Other steps in the design of SW filters are common to other filtration systems and include the determination of (i) the orifice diameter, (ii) the total height of the filter (heights of support layers + media + reactive zone + supernatant). The total height of a SW unit dictates the container to be used or

manufactured. In essence, in case ideal containers are not affordable, filters could be designed with available ones following a similar approach as described herein. Mathematical equations are also available to modify filter design accordingly [74,75,121,154].

4.5.3. Filter design

A modular SW-based filter for F^- removal for a 10-persons-large household is sought. It is assumed that available water source contains 10 mg/L F^- . The filter should be gravity fed and requires only medium replacement once or twice per year for maintenance. The filtered water should not contain more than 1.5 mg/L F^- . The reasoning herein is based on a filter container with the following dimensions: 10 cm inner diameter (D) and 100 cm height (L). Of the total length a depth of 70 cm is reserved to the reactive zone (RZ or H_{rz}) (Fig. 2). The volume of the reactive zone is thus: $\pi \cdot (D^2/4) \cdot H = 3.14 \cdot (10)^2/4 \cdot 70 = 5,500 \text{ cm}^3$ or 5.5 L. The total volume of the filter is 7.9 L, 2.6 L (corresponding to 30 cm) is left for sand and gravel.

The research question is which mass of a steel wool should be mixed to sand to build the 5.5 L reactive zone? The mass (or volume) of sand and SW vary but their total volume is fixed. An additional consideration is that the diameter of the sand particles (d) should not exceed 1.0 mm ($d/D \leq 0.1$). Typically sand particles used in such filters have d-value lower than 5.0 mm ([40] and refs. cited therein), meaning that filters with $D = 5.0 \text{ cm}$ can be envisaged as well. The case of a 100 % SW reactive zone using as received or properly pretreated material is envisaged, but will not be discussed herein. It is sufficient to consider that because of their mechanical properties the pure SW-Hrz may need to be supported.

For spherical particles, it has been established that the most sustainable Fe^0 :sand volumetric ratio is 1:3 (1 volume Fe^0 and 3 volumes of sand) [75,76,121,159,160]. This rule cannot be applied to fibrous SW. Therefore, fixing the H_{rz} volume and measuring the SW mass to be balanced with sand to fill it, is a good starting point. In this effort, both the resulting mass of sand and SW are to be documented, such that in the middle term, when enough data is available, statistical analysis could be performed to check whether rules of thumb can be elaborated for the construction of SW Fe^0 filters. It is understood that each tested SW is characterized for its elemental composition, its intrinsic reactivity (e.g. k_{EDTA} value see Bhatke-K et al. [156]) and the results are made available to the scientific community, at least as supporting information (appended to any publication, including technical documents).

A good starting point is to test the suitability of available SW materials with a design close to the one presented by Biswas and Bose [145] (based on SW mass in 125 cm³ of sand) and using testing methods summarized in Bhatke-K et al. [156] (e.g. k_{EDTA}). Selected materials will then be tested in the modular SW filter while varying the mass of SW in the 5.5 L reactive zone to optimize the filter efficiency. The experiments should be performed for at least one year regardless whether a breakthrough is observed or not. A specific objective will be to characterize the long term behavior of SW (through F^-/Fe breakthrough and permeability loss). The results for such experiments would demonstrate whether the initial modular system is satisfactorily or over-designed. In the case the system is over-designed there are at least two ways to improve it: (i) using a smaller filter (e.g. $D = 5.0 \text{ cm}$) or (ii) reducing the height of the reactive zone.

Summarized the first step on the way to popular Fe^0 filters for F^- removal can be described as follows:

- (i) Select a SW-material for its intrinsic reactivity using available tools [156];
- (ii) Characterize the selected SW-material for its elemental composition,
- (iii) Test the suitability of selected SW materials at small scale with a design comparable to the one described by Biswas and Bose [145];
- (iv) Test relevant SW materials with the modular filter described in this section (or a well-described modification).

The results of the modular test will assist to establish a base F⁻ treatment facility at bench scale. The system could be composed of several tanks in series of which just some are filled with specially prepared SW fibers mixed with sand. A design figure is appended (Fig. 2). Once the optimal SW/sand mixture is found, it can be contained in a fabric bag that is well-packed to facilitate insertion into and removal from the tank.

4.5.4. Beyond steel wool filters

The present study is a continuation of a vast effort to fully exploit the ability of metallic iron (Fe⁰) for in-situ generation of adsorbing (and reducing) agents for contaminant removal and thus water treatment. In-situ generation of adsorbing agents circumvents many limitations associated with the use of synthetic and natural adsorbents like ferrihydrites and other iron oxides/hydroxides [151,161]. Relevant limitations include mass transfer concerns and occupation of reactive sites by co-contaminants. In a properly designed Fe⁰ filter, a relevant contaminant (e.g. F⁻) and co-contaminants (e.g. PO₄³⁻, SO₄²⁻) will be independently removed from the aqueous phase. In other words, the main question with Fe⁰ filters is how long and with which reaction kinetics will they produce sufficient corrosion products for efficient contaminant removal?

Previous work mostly based on the premise that Fe⁰ is a reducing agent has not adequately addressed this question. In addition linear relationships used to describe reaction kinetics are not relevant for iron corrosion processes under environmental conditions [88]. Recent investigations have recalled the stochastic nature of environmental aqueous iron corrosion [128,129]. This means that relevant data, obtained under well-defined conditions are needed in order to derive appropriate kinetic parameters for long term efficient Fe⁰ filters (Eq. 5, Section 6.3). The achievement of such data is essential as small size Fe⁰ materials are allowed to corrode under water saturation for the first time. This assertion is valid despite five centuries experience in using iron shavings and scrap iron for the recovery of metals in the mining industry (cementation) [162,163].

To fill this gap, the research envisaged herein should be advantageously extended to larger size Fe⁰ materials. For example a series of Fe⁰ wires and Fe⁰ nails with various diameters and similar intrinsic reactivity could be tested in long term experiments to gain insights into the size-dependent corrosion of Fe⁰ materials during water treatment. The first advantage of this would be to determine a material that could be used in lieu of steel wool for longer operation time without maintenance (e.g. 18 or 24 months).

5. Economic considerations

The costs for a SW-based water filtration system include: the material cost (including SW and filter container), construction cost, operation cost, monitoring cost and decommission cost [40]. Currently,

every SW material might be regarded as unique to its site (local market). Thus, the overall cost of resulting Fe^0 filters will be site specific. Additionally, the manufacture of filter containers using local material should be encouraged.

The most reliable cost estimation for Fe^0 filters has been presented by Gottinger et al. [81,82] after testing such filters at pilot scale for As and U removal. A comparison to adsorptive filtration reveals that Fe^0 /sand is economical with respect to media and treatment cost, with a satisfactory service life. Used Fe^0 filings can be obtained for less than \$ 1.50/kg and the cost of manufacturing this filter is comparable to slow sand filter. Their calculations yielded a total treatment cost of less than \$ 0.01/L for the modular treatment train, including filter installation, media, operation and maintenance [83].

6. Discussion

6.1. The major challenge

Dental fluorosis and other health problems caused by elevated fluoride in drinking water have been known for years now [14,17,18,44,55]. Efficient F^- removal by chemical methods alone (e.g. precipitation) is not possible because the typical F^- concentration in natural waters is generally too low (typically less than 30 mg L^{-1}) to result in super-saturation and subsequent precipitation ([24] and refs. cited therein). Accordingly, only physical methods (adsorption, size exclusion) or a combination of physical and chemical methods can enable defluoridation within WHO standards ($\leq 1.5 \text{ mg L}^{-1}$). Additionally, to be relevant for remote locations, any method should be easily used by a layman while being efficient for the long term and affordable [6,24,39,40,72,78,]. Such a method has yet not been universally agreed, with the main two reasons being (i) the active surface of any filtration media is often very quickly passivated by adsorbed solutes, rendering the theoretical removal capacity extremely difficult to predict/quantify (hereafter Problem 1) and (ii) the long-term kinetics of removal process (e.g. dissolution of calcite or apatite) is difficult to quantify (hereafter Problem 2). To maintain compliance with the WHO standards the fluoride content of the treated water needs to be monitored during the entire service life of the filter. Problem 1 (reduced adsorption capacity) is common to all conventional adsorbent materials which maximum surface typically available at the start of the operation and then depreciating as a function of time. For cases where adsorbents are generated in-situ (e.g. $\text{Fe}^0/\text{H}_2\text{O}$ system) the situation may be different because continuously generated adsorbents separately remove fluoride ions and concurrent species (Cl^- , HCO_3^- , NO_3^- , PO_4^{3-} , SO_4^{2-}). However, a system can be customized to selectively remove F^- and other species. The main challenge is to couple the kinetics of Fe^0 corrosion to the other relevant variables (water quality, water flow velocity, temperature). In another phrase, Problem 2 (the long-term kinetics of iron corrosion) is the major challenge of designing efficient Fe^0 filters.

6.2. Steel wool based Fe^0 filters as starting point

The present work has introduced a concept for a systematic investigation of defluoridation using Fe^0 materials, starting with locally available steel wool. Fe^0 filtration technologies also provide the additional benefit of co-removing numerous other undesirable species when present even in trace

quantities. In order to progress most efficiently, [research with](#) a standard Fe^0 reference material (e.g. Connelly-GPM, iPutec GmbH or Peerless Metals) is required with results always published in open-access journals. All other operative conditions can be changed (other Fe^0 materials including steel wool, column dimensions, Fe^0 particle size) in particular the water chemistry as soon as natural waters are used. To optimise result comparability, sorbent materials (Fe^0 and additives) should be characterized extensively. Relevant parameters for granular material include: grain size, height of the reactive zone, point of zero charge, relative mass of Fe^0 and additives (or volumetric proportion), specific surface area, sphericity [75,154]. Moreover, testing common pollutants in addition to fluoride in synthetic solutions (e.g. $\text{As}^{\text{III}}/\text{As}^{\text{V}}$, F^- and U^{VI}) while using comparable experimental designs will accelerate the identification of strengths and weaknesses of the Fe^0 filtration technology. It is hoped that such approach will be soon adopted worldwide to enable the exploitation of the potential of this still innovative technology. This includes the identification of its limitation on a scientific basis.

6.3. Filter design considerations

The real merit of long term laboratory and field experiments based on principles exposed herein will be the estimation of long-term kinetics of iron corrosion for practical Fe^0 materials (including steel wool and iron wire) under relevant field conditions. In fact, first order models have been typically used to model contaminant removal in $\text{Fe}^0/\text{H}_2\text{O}$ systems [33,164]. Admittedly this approach was acceptable because Fe^0 was considered a reducing agent for the pollutants of concern. Now that the reducing transformation concept has been proven as false [35,165], alternative approaches are needed. Results from corrosion science have shown that the function describing instantaneous rate of iron corrosion is expressed as follows [166]:

$$v(t) = v_p + (v_0 - v_p) * \exp(-q_0 * t) \quad (5)$$

where q_0 is a constant, v_0 is the short-term and v_p the long-term average rate of Fe^0 corrosion. The instantaneous rate drops exponentially from v_0 to v_p , $v_0 > v_p$. Consequently, $v_p \leq v(t) \leq v_0$. The form of Eq. 5 allows relating the instantaneous rate, $v(t)$ to the average rate (v_p). “ v_0 ” is associated to laboratory measurements while “ v_p ” can be directly associated with the field measurements. Accordingly, results of experimental results performed following the concept presented herein will enable the determination of $-q_0$ and thus v_p . v_0 can be obtained from short-term column experiments using for example diluted EDTA solutions [167]. These parameters will enable a better design of Fe^0 filters. As expected, these results will be available for the whole Fe^0 research community [72].

6.4. Steel wool filters for all

It has been demonstrated herein that whilst steel wool filters can become reactively exhausted (corroded to end member hydroxides) within relatively short timescales such as two to three months, due to their cheap cost and high availability world-wide they would be easy to replace often in the developing world. Because steel wool is simply the starting material for a hand built and locally sourced water filter, there is a real potential to foster local industry during filter construction, which includes: (i) designing and manufacturing plastic or glass filter containers, (ii) purchasing and conditioning Fe^0 materials (including steel wool) and additives, and (iii) collecting and recycling spent

(corroded) filters [121,126,160,168]. In summary, the demand for decentralised water treatment can be exploited to start a new area of self-reliance in the developing world.

7. Concluding Remarks

Recent investigations have determined Fe^0 -bearing packed bed filters as a promising technology for decentralised and small scale clean drinking water provision. Fe^0 -bearing filters, however, have not yet been tested for this purpose. However, (i) taking into account the various drawbacks and challenges identified in the Fe^0 filter research and (ii) using a novel methodical approach for filter design, great progress in Fe^0 filter application can be expected in the near future.

Steel wool (SW) is suggested herein as an ideal candidate material for use as a Fe^0 -bearing filter material for the removal of F^- from drinking water. It is suggested that the use of SW as filtration media could be particularly beneficial in the developing world because of the materials low-cost and widespread abundance. A systematic methodology for the empirical testing of the suitability of SW as a filter for F^- is reported, including methodology to determine the impact of hydrological changes in filter design to incorporate SW corrosion processes (volumetric expansion) and methods to determine the influence of different concentrations of anions (e.g. added as sodium salts), cations (e.g. added in the nitrate form) and humic substances in the permeate fluid (contaminated water). Given the fibrous nature of SW, the addition of a constant SW mass (e.g. 25 g) to a defined volume of sand (e.g. 250 cm^3) would enable the characterization of the effects of shape, size and grading of sands on the efficiency of combined SW-sand filtration systems.

There is still overwhelming evidence that governments, charities and non-governmental organizations cannot effectively provide safe drinking water to all world-wide. For example, the United Nations (UN) Millennium goal of halving the proportion of people without safe drinking water in 1990 by 2015 was achieved by only a very close margin, and, in some parts of the world there is evidence to suggest that clean drinking water provision is worse than in the past. Some 780 million people still do not have access to safe drinking water. The concept presented herein is directly aimed to empower people for self-reliance in drinking water provision. An effective Fe^0 water filter can be both constructed and managed by local communities in remote locations even in politically unstable and economically deprived conditions because construction and maintenance will be cheap and easy to implement. The scientific community is invited to catalyze the development of such self-reliant, community led water filtration systems in the future, which if achieved will significantly benefit the health of millions of people world-wide. Evidence presented herein suggests that Fe^0 filters can provide a key part in this revolution, particularly for fluoride removal, and can be scaled-up or down to meet the requirements of each community.

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Author Contributions

The text of this article was written by Chicgoua Noubactep, Richard Crane and Karoli Njau with contributions from Arnaud Igor Ndé-Tchoupé and Hezron T. Mwakabona. Arnaud Igor Ndé-Tchoupé conducted background research on “steel wool for water treatment” while Hezron T. Mwakabona summarized previous work on all other adsorbents.

Conflicts of Interest

The authors declare no conflict of interest.

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